

# Climate change effects on Antarctic benthos: a spatially explicit model approach

Luciana Torre<sup>1</sup> · Paulo C. Carmona Tabares<sup>2</sup> ·  
Fernando Momo<sup>3,4</sup> · João F. C. A. Meyer<sup>5</sup> ·  
Ricardo Sahade<sup>1</sup>

Received: 29 June 2016 / Accepted: 28 January 2017  
© Springer Science+Business Media Dordrecht 2017

**Abstract** The Antarctic Peninsula is one of the regions on the Earth with the clearest evidence of recent and fast air warming. This air temperature rise has caused massive glacier retreat leading to an increased influx of glacier meltwater which entails hydrological changes in coastal waters, increasing sediment input and ice-scouring impact regime. It has been hypothesized that an increase of sediment load due to glacier retreat resulted in a remarkable benthic community shift in Potter Cove, a small inlet of the South Shetland Islands. In order to test this hypothesis, we developed an explicit spatial model to explore the link between sedimentation and ice-scouring increase upon four of the most conspicuous benthic species. This is a valuable novel approach since disturbances are strongly dependent of the space. The model takes into account sediment and population dynamics with Lotka-Volterra competition, a sediment-dependent mortality term and a randomized ice-scouring biomass removal. With the developed algorithm, and using a MATLAB environment, numerical simulations for scenarios with different sedimentation and ice-impact rates were undertaken in order to evaluate the effect of this phenomenon on biological dynamics. Comparing simulation results with biological data, the model not only recreates the spatial community distribution pattern but also seems to be able to recreate the shifts in abundance

---

**Electronic supplementary material** The online version of this article (doi:10.1007/s10584-017-1915-2) contains supplementary material, which is available to authorized users.

---

✉ Luciana Torre  
torreluciana@gmail.com

<sup>1</sup> Marine Ecology, Instituto de Diversidad y Ecología Animal (CONICET-UNC), Universidad Nacional de Córdoba, Facultad de Ciencias Exactas Físicas y Naturales., Av. Vélez Sarsfield 299, Córdoba 5000, Argentina

<sup>2</sup> Universidad del Quindío, Quindío, Colombia

<sup>3</sup> Instituto de Ciencias, Universidad Nacional de General Sarmiento, Los Polvorines, Buenos Aires, Argentina

<sup>4</sup> INEDES, Universidad Nacional de Luján - CONICET, Luján, Argentina

<sup>5</sup> IMECC, UNICAMP, Campinas, Brazil

under sedimentation enhancement, pointing out its importance as a structuring factor of polar benthic communities. Considering the challenges of Antarctic field work, this model represents a powerful tool for assessing, understanding, and predicting the effects of climate change on threatened Antarctic coastal ecosystems.

## 1 Introduction

There is a growing concern for understanding how climate change will affect polar ecosystems (Peck 2005; Smetacek and Nicol 2005; Barnes and Peck 2008; Kennicutt et al. 2014, 2015). In Antarctica, scientific priority setting is particularly important given the substantial cost and challenges of conducting research in the region. In a huge and international effort to establish these priorities, the first Antarctic and South Ocean Science Horizon Scan have recently identified the key scientific questions that need to be addressed in the following years. Among these questions, there is a clear concern about species and food web vulnerabilities at different spatial and temporal scales, under multiple stressors and environmental change drivers (Kennicutt et al. 2014, 2015). The west Antarctic Peninsula (WAP) is a region on the Earth with one of the most accelerated air warming trends. In fact, in the last few decades, it has shown a significant ice-mass loss as a consequence of ice shelf collapses and about 87% of its marine glaciers have been retreating and thinning (Cook et al. 2005; Pritchard and Vaughan 2007; Turner et al. 2009, 2016). This massive ice-mass imbalance entails hydrographical consequences including ice-scour and sedimentation increase (Bers et al. 2012; Schloss et al. 2012; Barnes et al. 2014) that could be considered the main benthic physical disturbance factors at coastal polar regions (Smale and Barnes 2008; Barnes and Peck 2008).

Ice-scour is a phenomenon in which icebergs pushed by wind, waves, or sea currents scrape the sea bottom removing and killing most of benthic fauna following a bathymetric pattern that becomes less intense with depth. Ice-scour has been intensively studied as a key structural force acting on benthic communities at high latitudes and essentially it decreases species abundance, richness, and coverage at a local scale but enhances regional diversity (Gutt 2001; Gutt and Starmans 2001; Brown et al. 2004; Teixidó et al. 2004; Gutt 2004; Barnes and Souster 2011).

Together with ice action, most coastal environments, especially fjords, are also influenced by sedimentation caused by seasonal glacier runoff. An increase in sedimentation rates can negatively affect aquatic ecosystems at many different levels, for example, by suffocating filter feeders, or affecting primary production by reducing light penetration (Pakhomov et al. 2003; Thrush et al. 2004; Deregibus et al. 2016). Studies of sedimentation effects on polar coastal benthos have been conducted almost exclusively in the Arctic. These studies clearly point out that sediment discharge is an important structuring factor of benthic communities that generates horizontal zonation. Diversity and species richness decrease towards inner fjord and glacier fronts (Carney et al. 1999; Włodarska-Kowalczyk and Pearson 2004; Włodarska-Kowalczyk et al. 2005). However, in Antarctica, a different pattern has been described, showing higher diversities and a clear dominance of epibenthic suspension feeders even in soft bottoms closer to the glacier fronts. Indeed, WAP fjords can be considered benthic diversity hotspots (Sahade et al. 1998; Grange and Smith 2013) that are currently threatened by increased sedimentation rates and ice disturbance frequency caused by massive glacier retreat along the WAP (Cook et al. 2005; Smale and Barnes 2008; Grange and Smith 2013). We recently reported a sudden shift in an Antarctic fjord, probably related to glacier retreat and increased inorganic matter runoff. Our results also suggested that benthic ecosystems could

cope with sedimentation until thresholds limits that once surpassed important shifts could be triggered (Sahade et al. 2015). The changes observed in the benthic community were decline of the species populations most vulnerable to sediment (mostly solitary ascidians) and population increase of the most tolerant ones, (like the bivalve *Laternula elliptica* and the pennatulid *Malacobelemnoides daytoni*) (Philipp et al. 2011; Torre et al. 2012, 2014; Sahade et al. 2015), coinciding with the highest sedimentation rates recorded for the last 100 years (Monien et al. 2011).

Considering the inherent difficulties to experimentally test hypotheses in Antarctica, we used a modeling approach to explore this complex disturbance scenario and provide new insights on the underlying ecological processes. As it has been summarized by Turner et al. (2009), different Antarctic environmental changes have intensively been modeled at different time and spatial scales, either for current or geological climatic events. Nevertheless, the benthic system has barely been explored. For instance, there are only a few models exploring the combined effects of benthic species longevity under different ice-scouring intensities (Potthoff et al. 2005; Johst et al. 2006). We present here an explicitly spatial model that links ice-scour and sedimentation increase on population dynamics of four key species. Based on a simpler previous approximation that only considered competition between the same groups of species and differential sensitivities to ice and sediment impact (Momo et al. 2008), we developed a more complex, realistic, and spatially explicit model. This allows us to explore the benthic community structure along a bathymetric gradient exposed to different sediment and ice-scouring intensities. Finally, contrasting simulation results with real data, the model will also help to understand not only the importance of each factor in the spatial distribution of the considered species but also how ice and sediment changes could affect them.

## 2 Methods and models

### 2.1 Species considered

We selected four species that together represent approximately the 40% of the Potter Cove animal benthic community biomass: *Molgula pedunculata*, *Ascidia challengerii*, *M. daytoni*, and *L. elliptica*. These species have shown the most critical abundances and changes in distribution patterns in the last 20 years (Sahade et al. 1998; Sahade et al. 2015). All of them inhabit soft bottoms and are filter feeders, so they compete for substratum and food sources.

### 2.2 The model

We developed an explicit spatial model that includes population growth, dispersion, and Lotka-Volterra competition between the four species selected. The lethal effect of sedimentation is added as mortality rate. The basic dynamics is based on the model of Momo et al. (2008). The ice-scouring effect was temporally and spatially randomized within a given impacted area. A detailed formulation of the model is explained in Supplementary Material 1.

### 2.3 Parameters

Most parameters were inferred from empirical data that was taken from multiple expeditions at Carlini station since 1993. Most of the data have been already published. Parameters and its

original sources are summarized in Table 1. Detailed calculation of parameters is explained in Supplementary Material 2.

## 2.4 Resolution method

To solve the equation system (3), a spatial discretization was performed with centered finite differences, and for time discretization, we used the Crank-Nicholson method (Strikwerda 2004; Leveque 2007; Burden and Faires 2010). As a result (after discretization), we found a nonlinear algebraic system that was solved by a predictor-corrector method. This method, originally implemented for solving parabolic problems with finite elements, is also adaptable for use with finite differences. Basically, the method consists of predicting the outcome by adjusting the system using nonlinear iterations of the algebraic system (Douglas et al. 1979).

## 2.5 Graphic representation

A two-dimensional spatial resolution is being analyzed where different depths are related to its distance to the coastline. We chose an area in Potter Cove which corresponds to the location of station 1 (E1) in the studies of Sahade et al. (1998, 2015) on benthic community structure changes over the years. This station also matches the hydrologic station of the inner cove, where salinity, MPS, chlorophyll, and temperature have been continuously recorded since 1991 (Bers et al. 2012; Schloss et al. 2012). This station is located on the south shore of the inner region of the cove, and it gets most glacier meltwater discharge through the two major runoff creeks (Fig. 1).

## 2.6 Simulated scenarios

Three sediment intensities were used: 70, 140, and 280 kg of sediment discharge distributed weekly and corresponding to S+, S++, and S+++, respectively. In the absence of accurate data about weekly or daily sediment loading in Potter Cove, these amounts were chosen in order to mimic water column sediment concentrations similar to those recorded between 1991 and 2009 (Schloss et al. 2012).

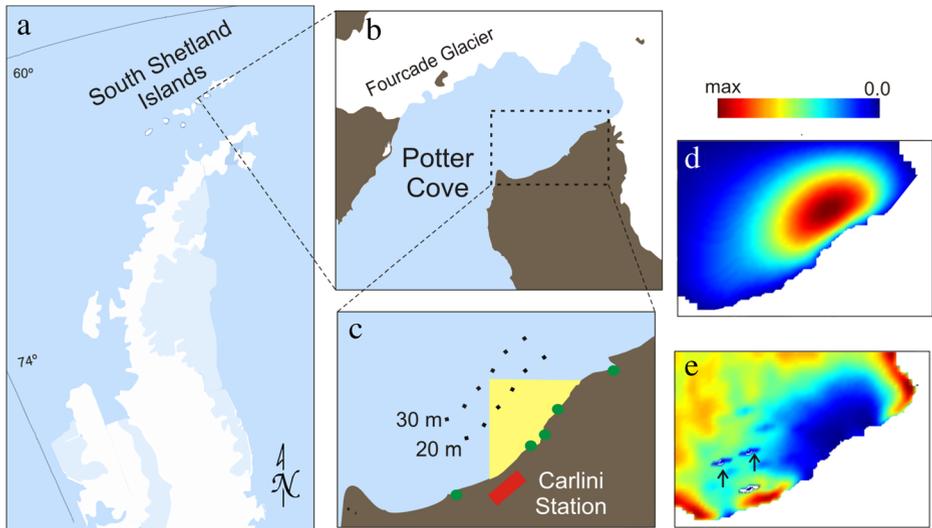
Three ice-scouring intensities were tested: 50, 200, and 400 impact events, corresponding to I+, I++, and I+++, respectively. As there is no data on ice impact for this area, many numbers of impacts were tested and the ones chosen were arbitrary, chosen to represent low, intermediate, and high impact levels. There were no differences in the final output over 400 impacts.

To compare the effect of both factors (ice and sediment) together, we evaluated many combinations of them. Nevertheless, for simplification and to compare with real data, we only consider analyzing those conditions reported for the area in the last 20 years. We had consistent data about sedimentation increase (Bers et al. 2012; Schloss et al. 2012; Sahade et al. 2015), but there was a lack of ice-scouring records for the area. However, bare substratum distribution was similar to the one obtained with the low ice-impact intensity scenario (I+). Thus, to create realistic sediment load simulations, we evaluated them maintaining a low ice-impact condition (low ice-impact + low sedimentation (I+S+); low ice-impact + high sedimentation (I+S+++)). An extra intermediate condition low ice-impact + medium sedimentation (I+S++) and its comparison with real data are detailed in Supplementary material 3.

**Table 1** Summary of model species parameters and its sources

Parameter	<i>M. pedunculata</i>	<i>A. challengeri</i>	<i>M. daytoni</i>	<i>L. elliptica</i>	Source
Average individual biomass (g AFDM)	2.23	2.75	0.07	4.05	Sahade (1999), Urban and Mercuri (1998), Servetto et al. (2013)
Maximum population biomass registered (g AFDM m <sup>-2</sup> )	60	20	0.5	80	Sahade (1999), Urban and Mercuri (1998), Servetto et al. (2013), Servetto and Sahade (2016)
Carrying capacity (g AFDM m <sup>-2</sup> )	85				Sahade et al. 1998, 2008, 2015
Dispersion coefficient km <sup>2</sup> w <sup>-1</sup>	1 × 10 <sup>-5</sup>	2 × 10 <sup>-5</sup>	1 × 10 <sup>-4</sup>	1 × 10 <sup>-4</sup>	This work
Intrinsic growth rate (year <sup>-1</sup> )	0.512	0.395	0.26	0.129	Rauschert (1991) Kühne (1997) Kowalke et al. (2001), Sahade et al. 2008, 2015 <sup>a</sup> , Urban and Mercuri
Competition coefficient					Sahade et al. 1998 <sup>a</sup>
	$\beta_{MpAc} = 1.868$	$\beta_{AsMp} = 2.079$	$\beta_{MdMp} = 1.967$	$\beta_{LeMp} = 1.972$	
	$\beta_{MpMd} = 1.760$	$\beta_{AcMd} = 1.862$	$\beta_{MdAc} = 1.620$	$\beta_{LeAc} = 1.556$	
	$\beta_{MpLe} = 1.129$	$\beta_{AsLe} = 1.146$	$\beta_{MdLe} = 1.142$	$\beta_{LeMd} = 1.781$	
Sediment mortality coefficient	$5.53 \times 10^{-2}$	$5.57 \times 10^{-3}$	–	$2.7 \times 10^{-4}$	Torre et al. 2012, Philipp et al. 2011 <sup>a</sup>

<sup>a</sup> Calculation of these parameters has been made as part of this work based on this previous work data



**Fig. 1** Location of the studied area in the Antarctic Peninsula (a). Map of Potter Cove with Fourcade Glacier (white) and ice-free areas (darker brown) (b). Closer view of the simulated area, green dots are pointing the main sediment sources (meltwater creeks); yellow marked correspond to the “ice-impacted zone” that mimics bathymetric ice-impact probability gradient. Transects of 20 and 30 m depth are also showed as pointed lines (c). Sediment simulated plume distribution example (d). Species-simulated distribution under sedimentation and ice-scouring example (arrows point ice-scouring event of biomass removal) (e). Colorimetric scale is auto-scaled from minimum (dark blue) to maximum (dark red) in each graph

## 2.7 Simulation output analysis

At the end of each simulation under different intensities of disturbance (ice, sediment, or both), we analyzed the resulting species distribution pattern. The graphs generated by the program were colored from light blue (minimum) to dark red (maximum) abundance values, but each graph is auto-scaled; therefore, the colors have only a relative meaning and they are not comparable across species. Therefore, to study the effect of different disturbance intensities or between species, we needed to directly analyze biomass values at the same selected points of each species grid. In order to get an output comparable with the photo-transects performed by Sahade et al. (2015), we selected points aligned parallel to the coastline representing the different transect depths (Fig. 1b). We compared real data with predictions of the model through nonparametric correlation Spearman analyzed with InfoStat version 2016 (Di Rienzo et al. 2016).

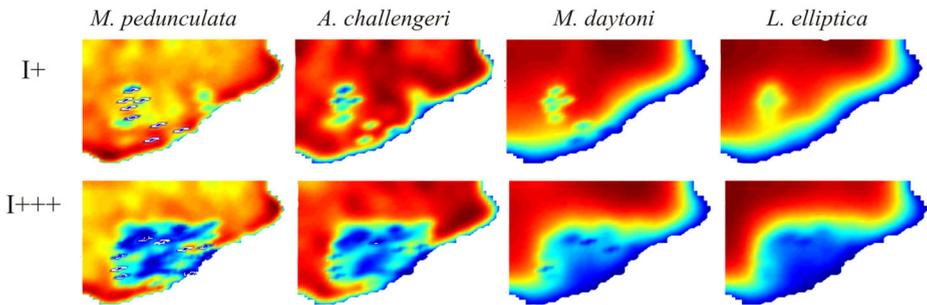
## 3 Results

### 3.1 Ice-scouring impact and sedimentation effect on species distribution pattern

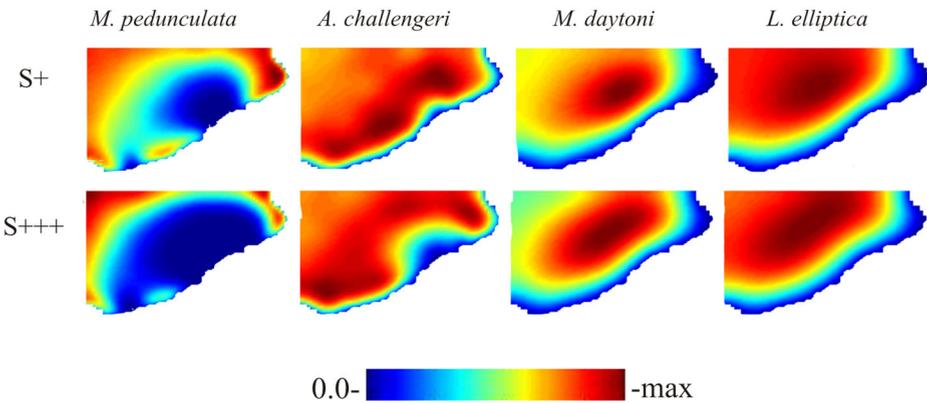
Removal of biomass by ice-impact did generate a relatively similar distribution pattern for all the simulated species, all showing their lower abundance in the near shore zone. As expected, the effect on each population did not exceed the impact area (Fig. 2a).

Simulated sediment discharge triggered a completely different distribution pattern between species. *M. pedunculata* was more abundant in deeper areas while *A. challengerii* seemed to be

## a Ice-scouring



## b Sedimentation

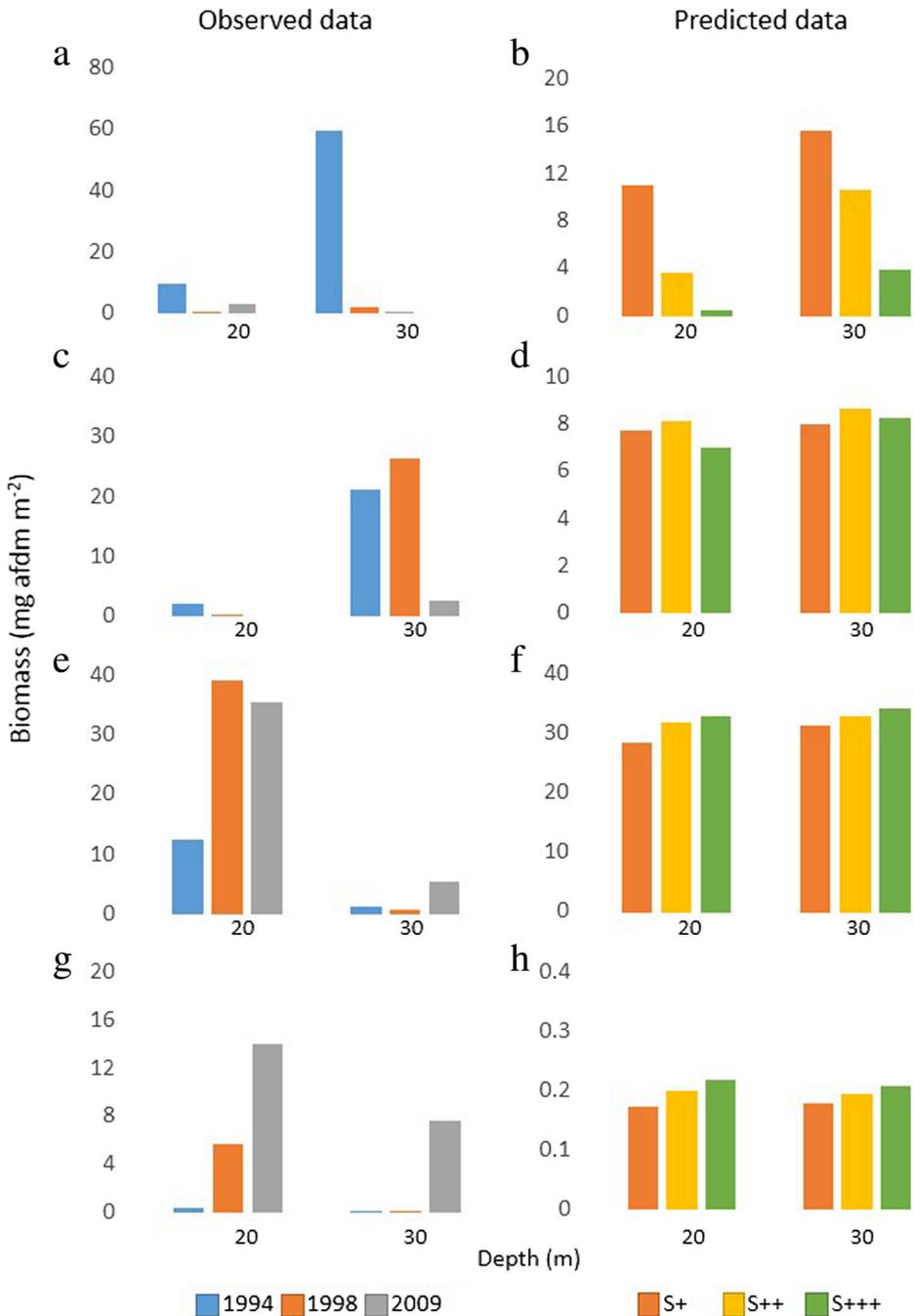


**Fig. 2** Final simulated distribution pattern of *M. pedunculata*, *A. challengeri*, *M. daytoni*, and *L. elliptica* under two ice-scouring impact conditions, low (I+) and high (I+++) (a) and under two sedimentary loads condition, low (S+) and high (S+++) (b). The colorimetric scale is auto-scaled from minimum (dark blue) to maximum (dark red) in each graph

more abundant at intermediate depths. Clearly, under higher sedimentation input, both retreated their distribution depths. On the contrary, *M. daytoni* and *L. elliptica* were both more abundant at shallower areas. The increase of sediment input showed an expansion of their distribution depth pattern (Fig. 2b). This apparently “positive” effect of sedimentation over the population of the last two species was actually an indirect consequence of the negative effect of sedimentation over their competitors (the ascidians). In the absence of a competitor, the distribution pattern of both species was completely different with almost no specific pattern or change at increasing sediment loads (data not shown).

### 3.2 Combined effect of ice-scouring and sedimentation against real data

The simulated sedimentation increase resulted in a decrease in the abundance of *M. pedunculata* changing their distribution pattern, almost disappearing at 20 m depths under the highest simulated concentration (Fig. 3b). This change had a very similar result to that one observed in Potter Cove for this species in the last 15 years ( $R^2 = 0.66$ ) (Fig. 3a). Although changes in *A. challengeri* biomass were not as dramatic as in *M. pedunculata*, they were well correlated with real data ( $R^2 = 0.71$ ). We could see a slight biomass increase from low to



**Fig. 3** Comparison between observed data in 1994, 1998, and 2009 surveys (Sahade et al. 1998, 2015) and modeled predicted data of three scenarios of low ice-impact (I+) and three sedimentary loads condition (S+, S++, S+++)) on distribution pattern and abundance of *M. pedunculata* (a, b), *A. challengeri* (c, d), *L. elliptica* (e, f), and *M. daytoni* (g, h) at 20 and 30 m depth

intermediate concentration to finally decrease at the higher one (Fig. 3d). This pattern of change coincided with observations in Potter Cove where, at 30 m depth, a biomass increase was recorded from 1994 to 1998 and finally decreased to a value below initial ones in 2009 (Fig. 3c). Both *L. elliptica* and *M. daytoni* showed biomasses increase proportional to sediment input increasing at every depth (Fig. 3f, h, respectively). This pattern coincided with the observed biomass change of these two species in the studied area (Fig. 3e, g, respectively) but was only significantly correlated for *M. daytoni* ( $R^2 = 0.77$ ).

## 4 Discussion

Antarctic benthic communities have been considered good sentinels for detecting environmental changes due to their relative stability and low velocities in biological processes. Although it was previously predicted, just recently, a sudden shift in a benthic community that could be related to the ongoing climate change has effectively been reported (Sahade et al. 2015), despite ice-scouring previously being considered the main factor structuring the Potter Cove benthic community (Sahade et al. 1998). Considering the physiological responses to sediments of the main species affected, and in the absence of other potential driving factor data, authors pointed to sedimentation increase as the possible causal factor. The results of this model support this idea.

### 4.1 Ice-scouring impact and sedimentation effect on species distribution pattern

It is widely known that Antarctic patchy distribution and depth zonation are mainly generated by ice biomass removal and substratum heterogeneity (Gutt and Starmans 2003; Gerdes et al. 2003; Smale 2008). Environments where organisms are grouped in patches generally show greater density variance than those homogeneously distributed (Margalef 2005). In our simulations, where all substratum is equally available for colonization, the decreasing density variance along depths could be considered an indicator of a bathymetric pattern of patchiness generated by ice influence. The model did not reproduce this expected ice caused pattern, where ice has always been considered the most important factor of species vertical distribution in almost all coastal polar environments (Dayton et al. 1969, 1974; Sahade et al. 1998; Nonato et al. 2000; Smale et al. 2007). It is known that ice effect diminishes with depth (Dayton et al. 1969, 1974; Smale 2008) and that most Antarctic coastal communities have zonation patterns along the bathymetric gradient. It is then supposed that highly impacted areas would determine the success of a good colonizer (a fast growing, highly reproductive, and high dispersal capacity species), while at deeper or less impacted areas, biological interactions will determine the dominance of the best competitors (Gutt and Starmans 2001; Poulin et al. 2002; Teixidó et al. 2004; Gutt et al. 2011). However, our model output was not completely consistent with this well-supported paradigm and neither the community studied in Potter Cove where *M. pedunculata*, having a higher growth rate and largely considered a pioneer species (Kowalke et al. 2001; Teixidó et al. 2004; Potthoff et al. 2005; Barnes and Conlan 2007, 2014; Gutt et al. 2011; Fillingner et al. 2013; Gutt et al. 2013), does not succeed in these shallower highly ice-impacted areas. Instead, *L. elliptica*, which is the species with the slowest growth and turnover rates (Urban and Mercuri 1998), dominates this more disturbed area together with *M. daytoni*. The success of *L. elliptica* in shallower zones could be related to its

ability to burrow, avoiding ice removal and the faster growth rate of *M. pedunculata* would not be enough to cope with this disturbance frequency. Nevertheless, this probably also indicates a synergic effect of other factors such as sedimentation, since *L. elliptica* and *M. daytoni* are more tolerant to it than *M. pedunculata* and these shallower areas are also more affected by sediments (Philipp et al. 2011; Torre et al. 2012). Nevertheless, a model is always an over simplification of reality, and in this case, it recreates the only aspects we know for sure about the ice-scouring phenomenon, the randomness of its occurrence, and its bathymetric occurrence pattern (Barnes 2016). The expectation was that ice could, as it has been hypothesized, recreate zonation pattern of this species in Potter Cove (Sahade et al. 1998). These results point out that we probably need to know more about the ice-impact phenomenon as the knowledge we have is not enough to recreate its effects on the benthic communities or that there could be other factors and interactions that should be taken into account that may intensify its effects.

Simulated sedimentation regenerated the distribution pattern of the species considered at Potter Cove. While it was expected for the most sensitive species *A. challengerii* and *M. pedunculata*, the distribution pattern of the other two species, *L. elliptica* and *M. daytoni* was an unexpected output of the model. Although *L. elliptica* is also affected by sedimentation (in a much lower extent than ascidians), in the model, it dominates the shallower areas together with *M. daytoni*. The result makes clear that this species distribution pattern is the result of an indirect effect of sedimentation that strongly affects ascidians. Meaning that its dominance at shallower areas depends on its higher tolerance to sedimentation, and the absence (due to sediment impact) of their stronger competitors, the ascidians.

#### 4.2 Combined effect of ice-scouring and sedimentation against real data

In general, the simulated sediment increase recreates very well the form of the change suffered by these four species in the last 20 years, but they fail in adjusting or achieving biomass values against real data. Most of these differences are related to the complexity of a system with more than 50 species where only four are being considered, ignoring the competitive or trophic interaction that could be taking place in nature. One of the most relevant differences is on the predicted biomass of *M. daytoni*, which may be indicating other phenomena that could be taking place and the model is not considering like a surprising episodic recruitment (Dayton et al. 2016a, b). Nevertheless, the model supported our hypothesis that increased sedimentation could be the causal factor of the observed sudden shift in Potter Cove benthic communities. Although the model developed is quite simple, it reproduced very well the main changes observed in Potter Cove benthic communities between 1994 and 2009 based on dynamic competition of four species with different sediment sensitivities.

This model did not intend to simulate sediment burial, which is one of the possible effects of sedimentation over benthic fauna but mainly associated to landslides or other massive sedimentation events described in Antarctica (Dayton et al. 2016a, b; Slattery and Bockus 1997), it was not the case of the community we studied. A slumping event would have provoked first a diminish in flat-shaped organisms (the ones that get more easily covered), though by the contrary showed an increase in the early stages of community shift, but as sedimentation increased in the system and in the model, these flat form sensitive ascidians were also affected (i.e., *A. challengerii* 1998 at 30 m depth (Fig. 3c)) (Sahade et al. 2015).

The idea of sediment generating horizontal zonation patterns on marine benthos is well extended in Arctic and temperate fjords (Thrush et al. 2004). It has been proposed as a factor that negatively affects diversity and richness. Włodarska-Kowalczyk and Wesławski (2001)

reported that as a consequence of turbidity, a primary production diminish near the sediment discharge also affects benthic biomass by reducing system capacity ( $K$ ). More recently, the same group found that species that dominate the nearest sediment discharge areas are far from the typical opportunistic species expected in highly disturbed areas. Instead, these species have lecithotrophic larvae and low offspring but seems to be the best adapted to resist sediment disturbance (Włodarska-Kowalczyk et al. 2005). However, due to the evolutionary history of the Antarctic continent, sedimentation was not taken into account until recently. Indeed, the existence of complex epibenthic communities on soft bottoms were explained by the absence of important sedimentary events such as those experienced by Arctic and temperate coastal and shelf systems during deglaciations (Gili et al. 2006). Unlike the rest of the world's coastlines, including the Arctic, Antarctica has virtually no sediment discharge from inland rivers. In fact, there are almost no estuaries because virtually the entire coast is covered with ice, and the low sediment supply comes primarily from the slow erosion of glaciers (Clarke et al. 2004). In the present scenario, where the retreating glaciers of the Antarctic Peninsula (Cook et al. 2005) are changing hydrography and the sedimentation regime not only locally (Monien et al. 2011; Schloss et al. 2012), but also extending their influence over 100 km from the coast (Dierssen et al. 2002), the importance of understanding sedimentary effects is acquiring a high relevance. It has been described that sediment influence decreases with its distance to the discharge source, but the sensitivity of filter-feeding species may increase with this distance. This means that specimens of species inhabiting areas frequently affected by sedimentation would deal better with sediment impact than the ones inhabiting pristine environments (Lohrer et al. 2006). If this were the case of susceptible Antarctic species, it is likely that the effect in more remote areas will be even more severe than that observed at the tip of Antarctic Peninsula, even at lower concentrations.

One of the biggest challenges for modelers nowadays is to enhance quality, resolution, and prediction capacity of global and regional models (Refsgaard et al. 2013; Boyd et al. 2014). However, to reach that goal, studies developed at local level, like the case of Potter Cove, are useful, since there is an extensive data set to contrast and test model outputs allowing an efficient feedback among models output, experimental work, and biological observations. These small but more detailed models will then point out which variables or interactions we should consider to enhance the quality of the regional ones.

**Acknowledgements** We thank the Carlini (former Jubany)–Dallmann staff for their support. Funding: Logistic and financial support were provided by Instituto Antártico Argentino, CONICET, FONCyT (Fondo para la Investigación Científica y Tecnológica), SECyT-UNC (Secretaría de Ciencia y Tecnología–UNC), DFG (Deutsche Forschungsgemeinschaft), and EU (European Union) via grants PICTO-DNA N° 119 and 36323, DFG project no. BR 775/25-1, IMCOAST (impact of climate induced glacial melting on marine coastal systems in the western Antarctic Peninsula region), IMCONet [FP7 IRSES (International Research Staff Exchange Scheme), action no. 319718], and PADI Foundation Grant 11234.

**Author attribution** LT: developed the model and parameters, carried out analyses, and wrote the original manuscript

PCCT: developed the model program and writing

FM: developed the model, discussed the results and writing.

JFCAM: developed the model program and writing.

RS: designed the project discussed the results and writing.

All authors discussed the manuscript.

### Compliance with ethical standards

**Conflict of interest** The authors declare that they have no conflict of interest.

## References

- Barnes DKA (2016) Iceberg killing fields limit huge potential for benthic blue carbon in Antarctic shallows. *Glob Chang Biol*. doi:[10.1111/gcb.13523](https://doi.org/10.1111/gcb.13523)
- Barnes DKA, Conlan KE (2007) Disturbance, colonization and development of Antarctic benthic communities disturbance, colonization and development of Antarctic benthic communities. *Philos Trans R Soc of LondB: Biol Sci* 362(1477):11–38. doi:[10.1098/rstb.2006.1951](https://doi.org/10.1098/rstb.2006.1951)
- Barnes DKA, Peck LS (2008) Vulnerability of Antarctic shelf biodiversity to predicted regional warming. *Clim Res* 37:149–163. doi:[10.3354/cr00760](https://doi.org/10.3354/cr00760)
- Barnes DKA, Souster T (2011) Reduced survival of Antarctic benthos linked to climate-induced iceberg scouring. *Nat Clim Chang* 1:365–368. doi:[10.1038/nclimate1232](https://doi.org/10.1038/nclimate1232)
- Barnes DKA, Fenton M, Cordingley A (2014) Climate-linked iceberg activity massively reduces spatial competition in Antarctic shallow waters. *Curr Biol* 24:R553–4. doi:[10.1016/j.cub.2014.04.040](https://doi.org/10.1016/j.cub.2014.04.040)
- Bers AV, Momo F, Schloss IR, Abele D (2012) Analysis of trends and sudden changes in long-term environmental data from King George Island (Antarctica): relationships between global climatic oscillations and local system response. *Clim Change* 116:789–803. doi:[10.1007/s10584-012-0523-4](https://doi.org/10.1007/s10584-012-0523-4)
- Boyd PW, Lennartz ST, Glover DM, Doney SC (2014) Biological ramifications of climate-change-mediated oceanic multi-stressors. *Nat Clim Chang* 5:71–79. doi:[10.1038/nclimate2441](https://doi.org/10.1038/nclimate2441)
- Brown KM, Fraser KPP, Barnes DKA, Peck LS (2004) Links between the structure of an Antarctic shallow-water community and ice-scour frequency. *Oecologia* 141(1):121–129. doi:[10.1007/s00442-004-1648-6](https://doi.org/10.1007/s00442-004-1648-6)
- Burden RL, Faires JD (2010) Numerical analysis, 9 edn. Brooks Cole
- Carney D, Oliver S, Armstrong C (1999) Sedimentation and composition of wall communities in Alaskan fjords. *Polar Biol* 22(1):38–49
- Clarke A, Aronson RB, Crame JA et al (2004) Evolution and diversity of the benthic fauna of the Southern Ocean continental shelf. *Antarct Sci* 16:559–568. doi:[10.1017/S0954102004002329](https://doi.org/10.1017/S0954102004002329)
- Cook AJ, Fox AJ, Vaughan DG, Ferrigno JG (2005) Retreating glacier fronts on the Antarctic Peninsula over the past half-century. *Science* 308:541–4. doi:[10.1126/science.1104235](https://doi.org/10.1126/science.1104235)
- Dayton PK, Robilliard GA, Devries AL (1969) Anchor ice formation in McMurdo Sound, Antarctica, and its biological effects. *Science* 163(3864):273–274
- Dayton PK, Robilliard GA, Paine RT, Dayton LB (1974) Biological accommodation in the benthic community at McMurdo Sound, Antarctica. *Ecol Monogr* 44(1):105–128. doi:[10.2307/1942321](https://doi.org/10.2307/1942321)
- Dayton PK, Hammerstrom K, Jarrell S et al (2016a) Unusual coastal flood impacts in Salmon Valley, McMurdo Sound, Antarctica. *Antarct Sci* 28(4):269–275. doi:[10.1017/S0954102016000171](https://doi.org/10.1017/S0954102016000171)
- Dayton PK, Jarrell S, Kim S et al (2016b) Surprising episodic recruitment and growth of Antarctic sponges: implications for ecological resilience. *J Exp Mar Biol Ecol* 482:38–55
- Deregisbus D, Quartino ML, Campana GL, Momo FR (2016) Photosynthetic light requirements and vertical distribution of macroalgae in newly ice-free areas in Potter Cove, South Shetland Islands, Antarctica. *Polar Biol* 39:153–166. doi:[10.1007/s00300-015-1679-y](https://doi.org/10.1007/s00300-015-1679-y)
- Di Rienzo JA, Balzarini M, Casanoves F et al (2016) InfoStat version 2016. <http://www.infostat.com.ar/>
- Dierssen HM, Smith RC, Vernet M (2002) Glacial meltwater dynamics in coastal waters west of the Antarctic peninsula. *Proc Natl Acad Sci U S A* 99:1790–5. doi:[10.1073/pnas.032206999](https://doi.org/10.1073/pnas.032206999)
- Douglas J, Dupont T, Ewing RE (1979) Incomplete iteration for time-stepping a Galerkin method for a Quasilinear parabolic problem. *SIAM J Numer Anal* 16:503–522. doi:[10.1137/0716039](https://doi.org/10.1137/0716039)
- Fillinger L, Janussen D, Lundälv T, Love S (2013) Report rapid glass sponge expansion after climate-induced antarctic ice shelf collapse. *Curr Biol* 23(14):1330–1334. doi:[10.1016/j.cub.2013.05.051](https://doi.org/10.1016/j.cub.2013.05.051)
- Gerdes D, Hilbig B, Montiel A (2003) Impact of iceberg scouring on macrobenthic communities in the high-Antarctic Weddell Sea. *Polar Biol* 26(5):295–301. doi:[10.1007/s00300-003-0484-1](https://doi.org/10.1007/s00300-003-0484-1)
- Gili J, Amtz WE, Palanques A et al (2006) A unique assemblage of epibenthic sessile suspension feeders with archaic features in the high-Antarctic. *Deep Sea Res II Top Stud Oceanogr* 53:1029–1052. doi:[10.1016/j.dsr2.2005.10.021](https://doi.org/10.1016/j.dsr2.2005.10.021)
- Grange LJ, Smith CR (2013) Megafaunal communities in rapidly warming fjords along the West Antarctic Peninsula: hotspots of abundance and beta diversity. *PLoS One* 8:e77917. doi:[10.1371/journal.pone.0077917](https://doi.org/10.1371/journal.pone.0077917)
- Gutt J (2001) On the direct impact of ice on marine benthic communities, a review. In *Ecological Studies in the Antarctic Sea Ice Zone* (pp. 157–168). Springer Berlin Heidelberg. doi:[10.1007/978-3-642-59419-9\\_21](https://doi.org/10.1007/978-3-642-59419-9_21)
- Gutt J (2004) Some “driving forces” structuring communities of the sublittoral Antarctic macrobenthos. *Antarct Sci*. doi:[10.1017/S0954102000000365](https://doi.org/10.1017/S0954102000000365)
- Gutt J, Starmans A (2001) Quantification of iceberg impact and benthic recolonisation patterns in the Weddell Sea (Antarctica). *Polar Biol* 24:615–619. doi:[10.1007/s003000100263](https://doi.org/10.1007/s003000100263)

- Gutt J, Starmans A (2003) Patchiness of the megabenthos at small scales: ecological conclusions by examples from polar shelves. *Polar Biol* 26(4):276–278. doi:10.1007/s00300-002-0468-6
- Gutt J, Barratt I, Domack E et al (2011) Biodiversity change after climate-induced ice-shelf collapse in the Antarctic. *Deep Sea Res Part II Top Stud Oceanogr* 58:74–83. doi:10.1016/j.dsr2.2010.05.024
- Gutt J, Cape M, Dimmler W et al (2013) Shifts in Antarctic megabenthic structure after ice-shelf disintegration in the Larsen area east of the Antarctic Peninsula. *Polar Biol* 36(6):895–906. doi:10.1007/s00300-013-1315-7
- Johst K, Gutt J, Wissel C, Grimm V (2006) Diversity and disturbances in the Antarctic megabenthos: feasible versus theoretical disturbance ranges. *Ecosystems* 9:1145–1155. doi:10.1007/s10021-006-0054-9
- Kennicutt MC, Chown SL, Cassano JJ et al (2014) Polar research: six priorities for Antarctic science. *Nature* 512: 23–25. doi:10.1038/512023a
- Kennicutt MC, Chown SL, Cassano JJ et al (2015) A roadmap for Antarctic and Southern Ocean science for the next two decades and beyond. *Antarct Sci* 27:3–18. doi:10.1017/S0954102014000674
- Kowalke J, Tatián M, Sahade R, Arntz W (2001) Production and respiration of Antarctic ascidians. *Polar Biol* 24: 663–669. doi:10.1007/s003000100266
- Kühne S (1997) Solitary Ascidians in the Potter Cove (King George Island, Antarctica) their ecological role and population dynamics. *Berichte zur Polarforschung (Reports on Polar Research)* 252
- Leveque RJ (2007) Finite difference methods for ordinary and partial differential equations: steady-state and time-dependent problems. SIAM 98
- Lohrer A, Hewitt J, Thrush S (2006) Assessing far-field effects of terrigenous sediment loading in the coastal marine environment. *Mar Ecol Prog Ser* 315:13–18. doi:10.3354/meps315013
- Margalef R (2005) *Ecología*, 10th edn. Ediciones Omega, Barcelona
- Momo F, Sahade R, Tatián M (2008) Benthic animal communities of Potter Cove (King George Island, Antarctica): observed patterns and explanatory models. *Ber Polarforsch Meeresforsch* 571:162–167
- Monien P, Schnetger B, Brumsack H-J et al (2011) A geochemical record of late Holocene palaeoenvironmental changes at King George Island (maritime Antarctica). *Antarct Sci* 23:255–267. doi:10.1017/S095410201100006X
- Nonato EF, Brito TA, De Paiva PC et al (2000) Benthic megafauna of the nearshore zone of Martel Inlet (King George Island, South Shetland Islands, Antarctica): depth zonation and underwater observations. *Polar Biol* 23(8):580–588. doi:10.1007/s0030000000129
- Pakhomov EA, Fuentes VL, Schloss IR et al (2003) Beaching of the tunicate *Salpa thompsoni* at high levels of suspended particulate matter in the Southern Ocean. *Polar Biol* 26:427–431. doi:10.1007/s00300-003-0494-z
- Peck LS (2005) Prospects for surviving climate change in Antarctic aquatic species. *Front Zool* 2(1):1–9. doi:10.1186/1742-9994-2-9
- Philipp EER, Husmann G, Abele D (2011) The impact of sediment deposition and iceberg scour on the Antarctic soft shell clam *Laternula elliptica* at King George Island, Antarctica. *Antarct Sci* 23:127–138. doi:10.1017/S0954102010000970
- Potthoff M, Johst K, Gutt J (2005) How to survive as a pioneer species in the Antarctic benthos: minimum dispersal distance as a function of lifetime and disturbance. *Polar Biol* 29:543–551. doi:10.1007/s00300-005-0086-1
- Poulin E, Palma AT, Féral J (2002) Evolutionary versus ecological success in Antarctic benthic invertebrates. *Trends Ecol Evol* 17:218–222
- Pritchard HD, Vaughan DG (2007) Widespread acceleration of tidewater glaciers on the Antarctic Peninsula. *J Geophys Res* 112:F3. doi:10.1029/2006JF000597
- Rauschert M (1991) Faunistic investigations in the benthos of King George Island. *Reports on Polar Research*. 76
- Refsgaard JC, Madsen H, Andréassian V et al (2013) A framework for testing the ability of models to project climate change and its impacts. *Clim Change* 122:271–282. doi:10.1007/s10584-013-0990-2
- Sahade R (1999) Patrones y procesos en una comunidad Epi-Bentónica Antártica: el ejemplo de Caleta Potter. Universidad Nacional de Córdoba
- Sahade R, Tatián M, Kowalke J et al (1998) Benthic faunal associations on soft substrates at Potter Cove, King George Island, Antarctica. *Polar Biol* 19:85–91. doi:10.1007/s003000050218
- Sahade R, Tarantelli S, Tatián M, Mercuri G (2008) Benthic community shifts: A possible linkage to climate change? *Berichtezur Polar und Meeresforsch* 571:331–337
- Sahade R, Lagger C, Torre L et al (2015) Climate change and glacier retreat drive shifts in an Antarctic benthic ecosystem. *Sci Adv*. doi:10.1126/sciadv.1500050
- Schloss IR, Abele D, Moreau S et al (2012) Response of phytoplankton dynamics to 19-year (1991–2009) climate trends in Potter Cove (Antarctica). *J Mar Syst* 92:53–66. doi:10.1016/j.jmarsys.2011.10.006
- Servetto N, Sahade R (2016) Reproductive seasonality of the Antarctic Sea Pen *Malacobelemn daytoni* (Otocorallia, Pennatulacea, Kophobelemnidae). *PLoS One* 11(10):e0163152. doi:10.1371/journal.pone.0163152
- Servetto N, Torre L, Sahade R (2013) Reproductive biology of the Antarctic “sea pen” *Malacobelemn daytoni* (Otocorallia, Pennatulacea, Kophobelemnidae). *Polar Res* 32

- Slattery M, Bockus D (1997) Sedimentation in McMurdo Sound, Antarctica: a disturbance mechanism for benthic invertebrates. *Polar Biol* 18:172–179. doi:[10.1007/s003000050174](https://doi.org/10.1007/s003000050174)
- Smale D (2008) Continuous benthic community change along a depth gradient in Antarctic shallows: evidence of patchiness but not zonation. *Polar Biol* 31:189–198. doi:[10.1007/s00300-007-0346-3](https://doi.org/10.1007/s00300-007-0346-3)
- Smale D, Barnes DKA (2008) Likely responses of the Antarctic benthos to climate-related changes in physical disturbance during the 21st century, based primarily on evidence from the West Antarctic Peninsula region. *Ecography* 31(3):289–305. doi:[10.1111/j.0906-7590.2008.05456.x](https://doi.org/10.1111/j.0906-7590.2008.05456.x)
- Smale D, Barnes DKA, Fraser K (2007) The influence of depth, site exposure and season on the intensity of iceberg scouring in nearshore Antarctic waters. *Polar Biol* 30:769–779. doi:[10.1007/s00300-006-0236-0](https://doi.org/10.1007/s00300-006-0236-0)
- Smetacek V, Nicol S (2005) Polar ocean ecosystems in a changing world. *Nature* 437:362–8. doi:[10.1038/nature04161](https://doi.org/10.1038/nature04161)
- Strikwerda JC (2004) Finite difference schemes and partial differential equations. 2nd ed. SIAM
- Teixidó N, Garrabou J, Gutt J, Arntz WE (2004) Recovery in Antarctic benthos after iceberg disturbance: trends in benthic composition, abundance and growth forms. *Mar Ecol Prog Ser* 278:1–16. doi:[10.3354/meps278001](https://doi.org/10.3354/meps278001)
- Thrush SF, Hewitt JE, Cummings VJ et al (2004) Muddy waters: elevating sediment input to coastal and estuarine habitats. *Front Ecol Environ* 2:299–306. doi:[10.1890/1540-9295\(2004\)002\[0299:MWESIT\]2.0.CO;2](https://doi.org/10.1890/1540-9295(2004)002[0299:MWESIT]2.0.CO;2)
- Torre L, Servetto N, Eöry MLM et al (2012) Respiratory responses of three Antarctic ascidians and a sea pen to increased sediment concentrations. *Polar Biol* 35:1208. doi:[10.1007/s00300-012-1208-1](https://doi.org/10.1007/s00300-012-1208-1)
- Torre L, Abele D, Lagger C et al (2014) When shape matters: strategies of different Antarctic ascidians morphotypes to deal with sedimentation. *Mar Environ Res* 99:179–87. doi:[10.1016/j.marenvres.2014.05.014](https://doi.org/10.1016/j.marenvres.2014.05.014)
- Turner J, Bindschadler R, Convey P et al (2009) Antarctic climate change and the environment: a contribution to the International Polar Year 2007–2008. Scientific Committee on Antarctic Research, Cambridge
- Turner J, Lu H, White I et al (2016) Absence of 21st century warming on Antarctic Peninsula consistent with natural variability. *Nature* 535:411–415. doi:[10.1038/nature18645](https://doi.org/10.1038/nature18645)
- Urban HJ, Mercuri G (1998) Population dynamics of the bivalve *Latemula elliptica* from Potter Cove, King George Island, South Shetland Islands. *Antarct Sci* 10(02):153–160
- Wlodarska-Kowalczyk M, Pearson TH (2004) Soft-bottom macrobenthic faunal associations and factors affecting species distributions in an Arctic glacial fjord (Kongsfjord, Spitsbergen). *Polar Biol* 27:155–167. doi:[10.1007/s00300-003-0568-y](https://doi.org/10.1007/s00300-003-0568-y)
- Wlodarska-Kowalczyk M, Weslawski J (2001) Impact of climate warming on Arctic benthic biodiversity: a case study of two Arctic glacial bays. *Clim Res* 18:127–132
- Wlodarska-Kowalczyk M, Pearson T, Kendall M (2005) Benthic response to chronic natural physical disturbance by glacial sedimentation in an Arctic fjord. *Mar Ecol Prog Ser* 303:31–41. doi:[10.3354/meps303031](https://doi.org/10.3354/meps303031)